

Assessing contemporary methylmercury exposure to resident and migratory bird species of Belize

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Front cover photos: Secondary forest habitat present at the Toucan Ridge Ecology and Education Society (TREES, top), Ruddy Woodcreeper (*Dendrocincla homochroa*, bottom left), and Orange-breasted Falcon (*Falco deiroleucus*, bottom right). All photos in this report were taken by Christopher Sayers.

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Abstract

Anthropogenic mercury (Hg) is a systemic environmental toxin throughout the Neotropics that may pose population-level impacts to fish- and invertebrate-eating bird species. During a rapid avian Hg assessment conducted in the wet season of 2021 by the Biodiversity Research Institute and the Foundation for Wildlife Conservation, we found 72% of well-sampled species ($n \geq 10$) had lower mean Hg concentrations than in previous surveys conducted from 2007–2019, while 80% of sites had higher mean Hg concentrations. When excluding poorly-sampled species ($n < 10$) from our most recent survey, only the American Pygmy Kingfisher (*Chloroceryle aenea*) had blood Hg concentrations that exceeded adverse effect thresholds associated with reproductive impairment. Previous Neotropical bird Hg biomonitoring efforts (Sayers et al. *In review*) highlight central Belize as a biological Hg hotspot, but this contemporary assessment may suggest that Hg biomagnification is deescalating or atmospheric Hg deposition is decreasing throughout the Yucatán Peninsula. To expand our understanding of these evolving dynamics in Belize, as a function of BRI's on-the-ground involvement, several Belizean conservation organizations are actively contributing to bird tissue collection efforts in collaboration with the Tropical Research for Avian Conservation and Ecotoxicology (TRACE) Initiative.

Background

Mercury (Hg) is a widely-distributed and persistent environmental pollutant with detrimental implications for global health. International government organizations to curtail global Hg emissions, such as the United Nations Minamata Convention on Mercury, have made transformative progress in generating the research and legislation necessary to regulate polluting industries throughout the Global North. However, such action is severely lacking in tropical nations despite rising emissions throughout the Global South. As a function of rising gold demand and price since the beginning of the century, the rapid expansion of artisanal and small-scale gold mining (ASGM) has catapulted the global tropics as the leading contributor of environmental Hg pollution (UNEP 2019). This paradigm shift is cause for major concern in how we consider and assess Hg health risks to tropical ecosystems and their inhabitants.

Tropical wet forest biomes maintain biogeochemical conditions, such as elevated precipitation and seasonal flooding, that enhance ASGM emissions by promoting Hg bioavailability. Recent work by Sayers et al. (*In review*) and Gerson et al. (2022) have begun to pinpoint which biotic and abiotic factors predispose biodiversity to high Hg exposure — using birds as accessible and economical bioindicators. Tropical Hg biomagnification is expected to intensify for 1) sites within close proximity (≈ 2 km) to artisanal gold mining or processing zones, 2) sites with intact, robust forest canopies, 3) sites where Hg methylating habitats, such as wetlands, are abundant throughout the landscape, and 4) organisms feeding at high trophic

positions. To summarize, fish- and invertebrate-eating bird species foraging within primary rainforest or wetland habitats adjacent to ASG mining or processing sites are predicted to have exceedingly high blood and feather Hg concentrations relative to the remaining avian community. There is now mounting evidence associating high Hg exposure with reproductive, behavioral, immunological, neurological, and physiological impairment in wildlife (Depew et al. 2012; Dietz et al. 2013; Scheuhammer et al. 2015), particularly birds (Evers et al. 2005; Ackerman et al. 2016; Fuchsman et al. 2017; Evers 2018; Whitney and Cristol 2018). And while these negative effects are seldom applied at the population level, Hg exposure could be a “silent” contributor to avian biodiversity loss throughout the Americas (Rosenberg et al. 2019; Stouffer et al. 2020; Sherry 2021) — highlighting the need for consistent, widespread Hg monitoring.

Since 2007, the Biodiversity Research Institute (BRI) has collaborated with numerous conservation organizations in Belize to begin understanding Hg dynamics within terrestrial tropical ecosystems — using resident and migratory bird species as key bioindicators. Preliminary objectives included developing Hg exposure profiles for terrestrial bird species, documenting geographic Hg hotspots and important methylating habitats, and communicating findings to government agencies and local communities to raise awareness about the impacts of Hg on human, wildlife, and ecosystem health. After over a decade of infrequent avian Hg sampling expeditions and the analysis of over 450 blood and feather samples, central Belize has emerged as a biological Hg hotspot, which has produced elevated concentrations across a variety of avifauna (Evers and Burton 2020; Sayers et al. *In review*).

While previous work has defined a valuable baseline from which to compare future data, the data showcased in this report represent the most recent Hg assessment for central Belize and can help illustrate how avian Hg concentrations and risk have changed over the past 15 years.

Objectives

1. Generate contemporary avian Hg exposure profiles and risk assessments
2. Quantify how avian Hg exposure and risk are changing through time
3. Identify several relatively common, broadly distributed resident bird species to serve as future sentinel Hg bioindicators across Latin America
4. Create and strengthen scientific partnerships with bird conservation organizations by holding in-person meetings and lectures

Methods

Study areas

To collect avian tissue samples for later Hg analysis, we conducted ground-level mist net surveys from August 25 through October 15, 2021 at the Tropical Education Center, Runaway Creek Nature Reserve, Monkey Bay Wildlife Sanctuary, and Toucan Ridge Ecology and Education Society with primary assistance from staff at the Foundation for Wildlife Conservation, Tropical Education Center, 28 George Price Highway, P.O. Box 368, La Democracia, Belize District, Belize. We selected study locations in a variety of habitats including riparian scrub, pine savanna, and secondary forest. Site-specific capture and habitat information are featured in Appendix 1.



Figure 1. Karst hill broadleaf forest, pine savanna, and seasonal wetland matrix at Runaway Creek Nature Reserve (left) and mist net array along Sibun River (right).

Sample Collection

We captured birds using 12 m mist nets and identified individuals to an appropriate species, sex, age, and molt cycle using a local bird field guide (Fagan and Komar 2016) and either a calendar-based, or the preferred, Wolfe-Ryder-Pyle (WRP) cycle-based age-classification system (Wolfe et al. 2010; Johnson et al. 2011; Tórréz and Arendt 2012, 2017; Pyle et al. 2016) described for Neotropical bird families in Johnson and Wolfe (2017). Whenever possible, we banded all taxa, excluding hummingbirds (Trochilidae), with an aluminum leg band from the US Fish and Wildlife Service and the National Band & Tag Company (<https://www.nationalband.com/>, Newport, Kentucky, USA). In unique circumstances when we

did not possess aluminum leg bands in the proper size, we banded individuals with a unique combination of colored plastic legs bands and assigned the corresponding samples a unique identification number. We also measured and assessed birds for feather molt and wear, skull ossification, fat stores, muscle mass, wing chord, tarsus length, bill dimensions, tail length, body mass, and reproductive stage via cloacal protuberance and brood patch.

We followed tissue collection, preparation, and storage methods provided by the Biodiversity Research Institute (Evers et al. 2021). Whenever possible, we collected 30–60 μL of blood from the cutaneous ulnar vein using 75 μL heparinized capillary tubes, sealed tubes at both ends using Critocaps™ or Critoseal™, placed tubes into a plastic vacutainer, and stored samples in a cooler with ice packs. We transferred blood samples into a freezer within 8 hours of collection where they were stored below $-4\text{ }^{\circ}\text{C}$ until laboratory analysis. For feather sampling, we collected the two outermost tail feathers and 5–6 flank feathers and stored samples in paper coin envelopes at ambient temperature.

While we collected multiple tissue types from each individual to obtain a holistic sample, we only showcase the whole blood data in this report. Whole blood best represents recent dietary Hg exposure from days to weeks, which makes it an ideal sampling matrix for rapid site assessments (Evers et al. 2005; Evers 2018).



Figure 2. Preparing a Green Kingfisher (*Chloroceryle americana*) for blood sampling via brachial venipuncture.

Laboratory analysis

We performed total mercury analyses at the Biodiversity Research Institute Toxicology Lab (Portland, Maine, USA) using a thermal decomposition and atomic absorption spectrophotometry technique with a Nippon MA-3000 direct Hg analyzer. We assumed that nearly all total mercury (THg, > 95%) in whole blood (Rimmer et al. 2005; Edmonds et al. 2010) to be in the bioavailable, methylmercury (MeHg) form. Therefore, all sampled tissue concentrations represent MeHg contamination of Neotropical avifauna. We followed United States Environmental Protection Agency (EPA) SW-846 Method 7473, “Mercury in solids and solutions by thermal decomposition, amalgamation, and atomic absorption spectrophotometry” (USEPA 1998), and used quality control methods (including certified reference materials DOLT-5 and ERM-CE464) to ensure consistent analytical precision and accuracy. We excluded any sample that registered below the analyzer’s lower detection limit (0.05 ng THg) from the final database.

Results and Discussion

Hg exposure and risk

We present a collective database containing 546 whole blood THg samples from 7 orders, 27 families, and 86 bird species throughout central Belize collected from 2007–2021. The most well-sampled bird families included New World warblers (Parulidae; $n = 170$), tyrant flycatchers (Tyrannidae; $n = 64$), woodcreepers (Furnariidae; $n = 45$), tanagers (Thraupidae; $n = 44$), wrens (Troglodytidae; $n = 35$), manakins (Pipridae; $n = 32$), vireos (Vireonidae; $n = 28$), and kingfishers (Alcedinidae; $n = 25$).

Recent research on Neotropical bird Hg exposure documents consistently elevated Hg concentrations in bird communities of the Yucatán Peninsula from 2007–2019 and has highlighted central Belize as a biological Hg hotspot relative to other regions throughout the Neotropical realm (Sayers et al. *In review*). Belize may face elevated Hg deposition because of gaseous elemental Hg emissions from local landfill incineration, coal combustion in central Mexico (UNEP 2019), as well as industrial and artisanal gold mining in the Chiquibul/Maya Mountains (Cornec 2010; Briggs et al. 2013; Manzanero 2014; Rath 2016). In addition, there is an abundance of key Hg methylating habitats at some of the Belize sampling sites, such as

seasonal wetlands at Runaway Creek Nature Reserve, that we expect can readily convert inorganic emissions to a bioavailable form.

Our contemporary Hg assessment conducted in the wet season of 2021 opposes this present school of thought by providing generally positive, but paradoxical, results for Hg biogeochemical cycling in Belize. We found that 72% of well-sampled bird species ($n \geq 10$) had lower mean Hg concentrations relative to previous surveys conducted from 2007–2019 (Fig. 1), while 80% of sites had higher mean Hg concentrations (Fig. 2). Perhaps the most valuable way of assessing these results is to put them into context of established adverse effect thresholds. In our contemporary survey, the American Pygmy Kingfisher (*Chloroceryle aenea*) was the only species that exhibited blood Hg concentrations associated with a 10% loss in reproductive success (Fig. 3; Jackson et al. 2011). This is a marked improvement compared to historical surveys, in which a total of 4 species exceeded reproductive impairment thresholds (Fig. 3). Therefore, these factors may indeed suggest that avian Hg risk and biomagnification are deescalating or atmospheric Hg deposition is decreasing throughout the Yucatán Peninsula — uplifting, but unexpected findings that are deserving of further scrutiny.

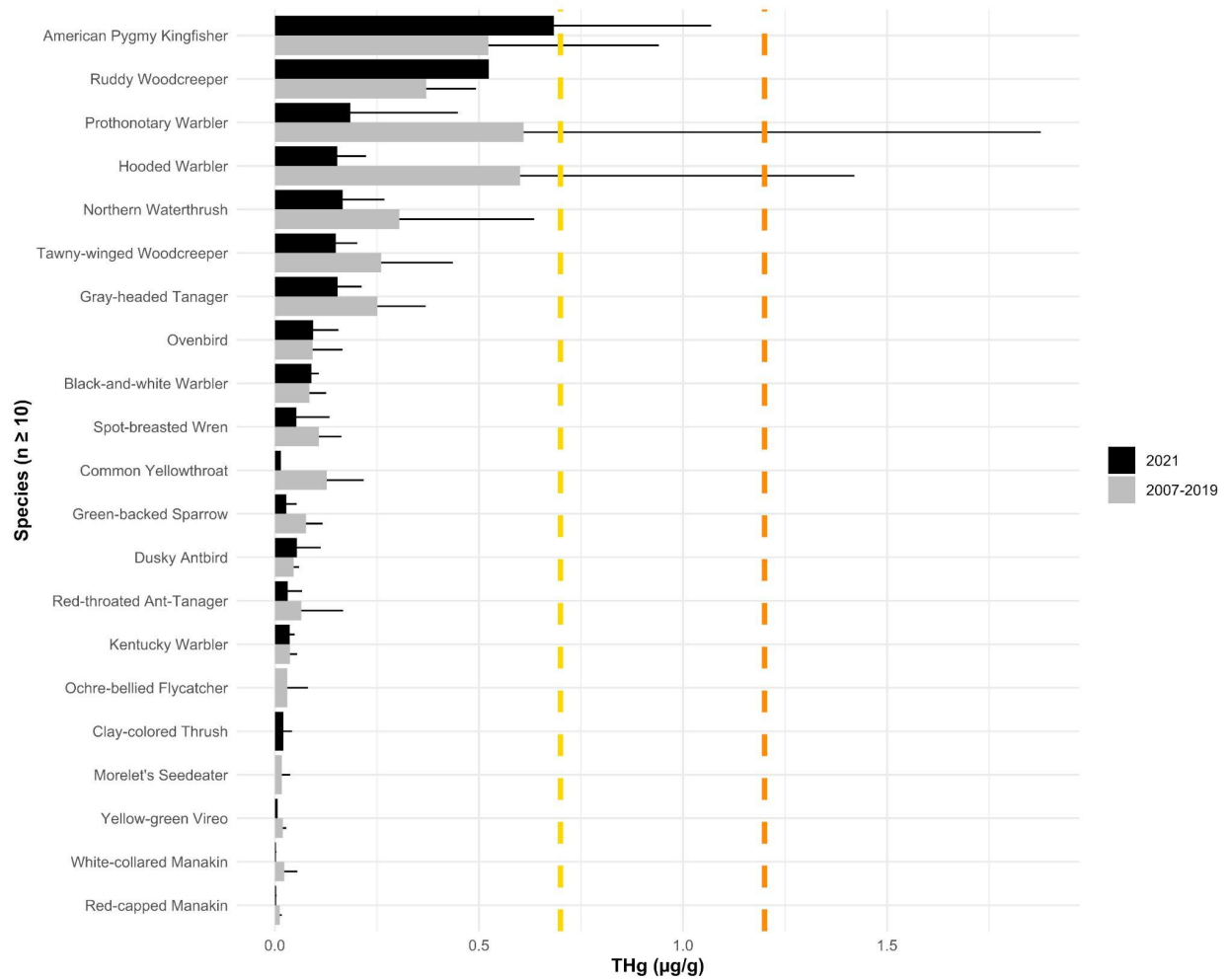


Figure 1. Arithmetic mean \pm standard deviation of whole blood total mercury (THg) concentrations ($\mu\text{g/g}$) among bird species sampled throughout central Belize from 2007–2021. The dashed horizontal lines indicate risk thresholds defined by Jackson et al. (2011) where $\geq 0.7 \mu\text{g/g}$ ww represents a $\geq 10\%$ decline in reproductive success (yellow) and $\geq 1.2 \mu\text{g/g}$ ww represents a $\geq 20\%$ decline in reproductive success (orange).

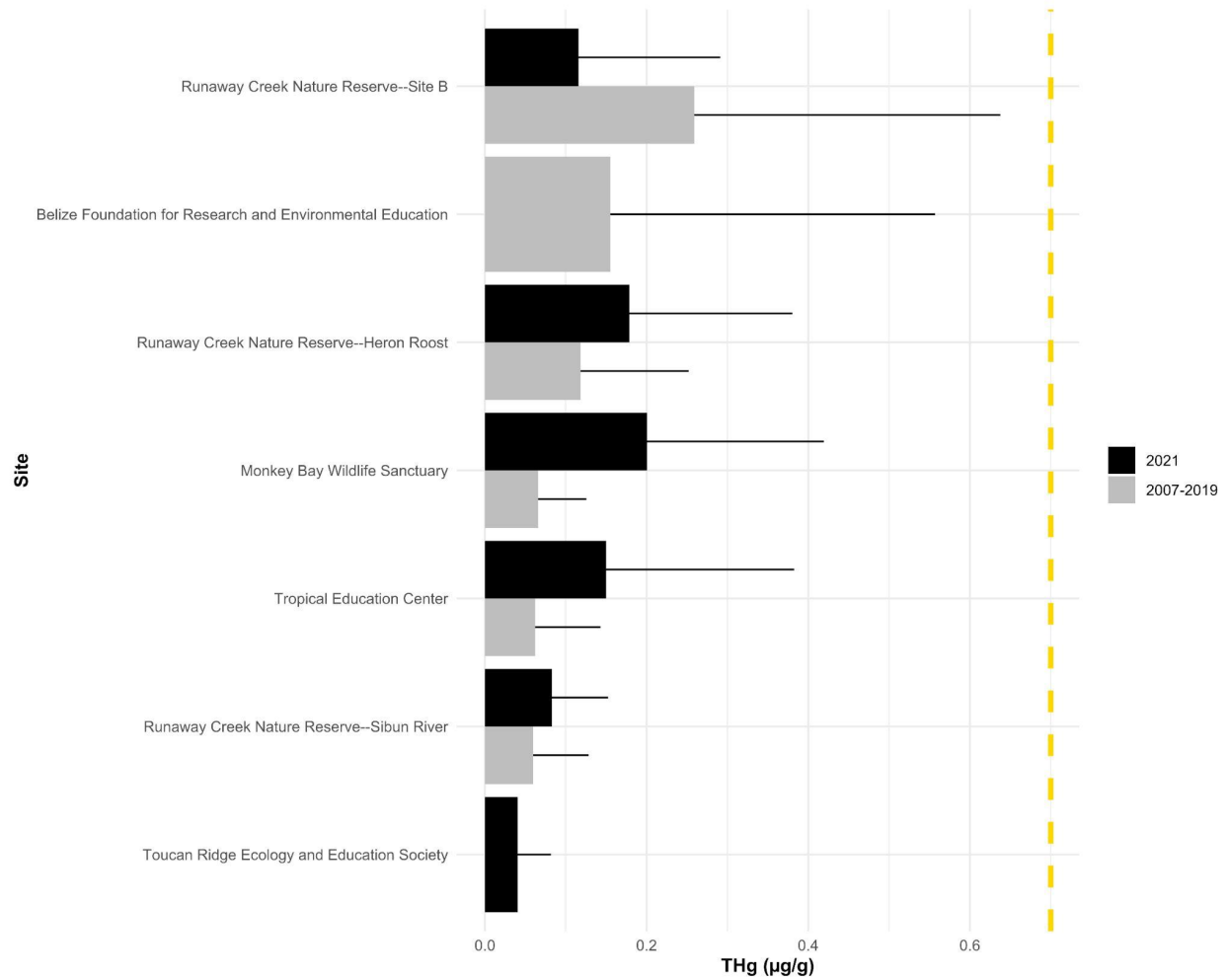


Figure 2. Arithmetic mean \pm standard deviation of whole blood total mercury (THg) concentrations ($\mu\text{g/g}$) among sites sampled throughout central Belize from 2007–2021. The yellow dashed horizontal line indicates a risk threshold defined by Jackson et al. (2011) where $\geq 0.7 \mu\text{g/g}$ ww represents a $\geq 10\%$ decline in reproductive success.

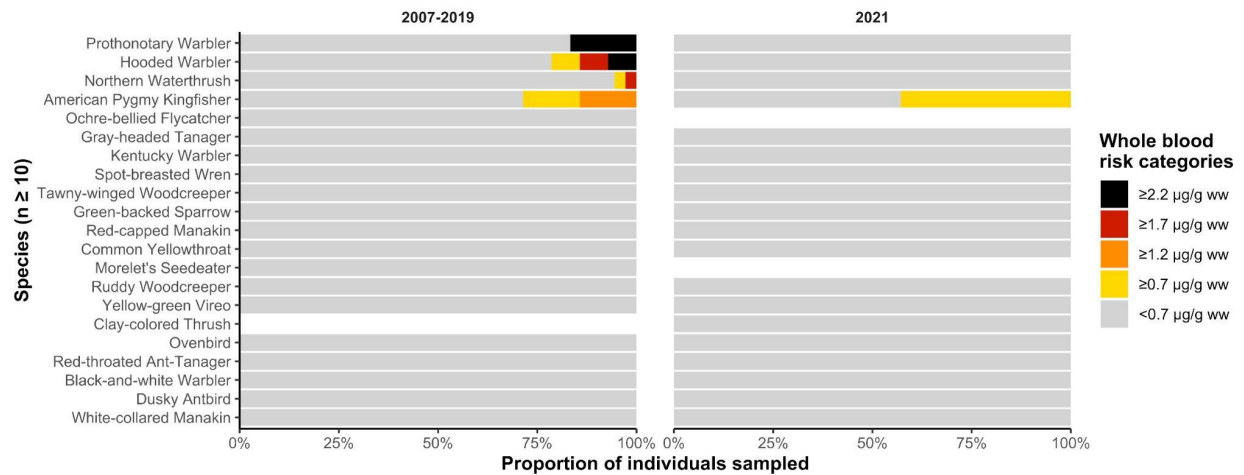


Figure 3. Proportion of Neotropical bird species sampled for whole blood throughout central Belize from 2007–2021 that may be subject to different degrees of reproductive failure via MeHg exposure. Species with fewer than 10 samples were excluded. Whole blood risk categories defined by Jackson et al. (2011) include: < 0.7 µg/g ww (gray, represents a ≤ 10% decline in reproductive success), ≥ 0.7 µg/g ww (yellow, represents a ≥ 10% decline in reproductive success), ≥ 1.2 µg/g ww (orange, represents a ≥ 20% decline in reproductive success), ≥ 1.7 µg/g ww (red, represents a ≥ 30% decline in reproductive success), and ≥ 2.2 µg/g ww (black, represents a ≥ 40% decline in reproductive success).

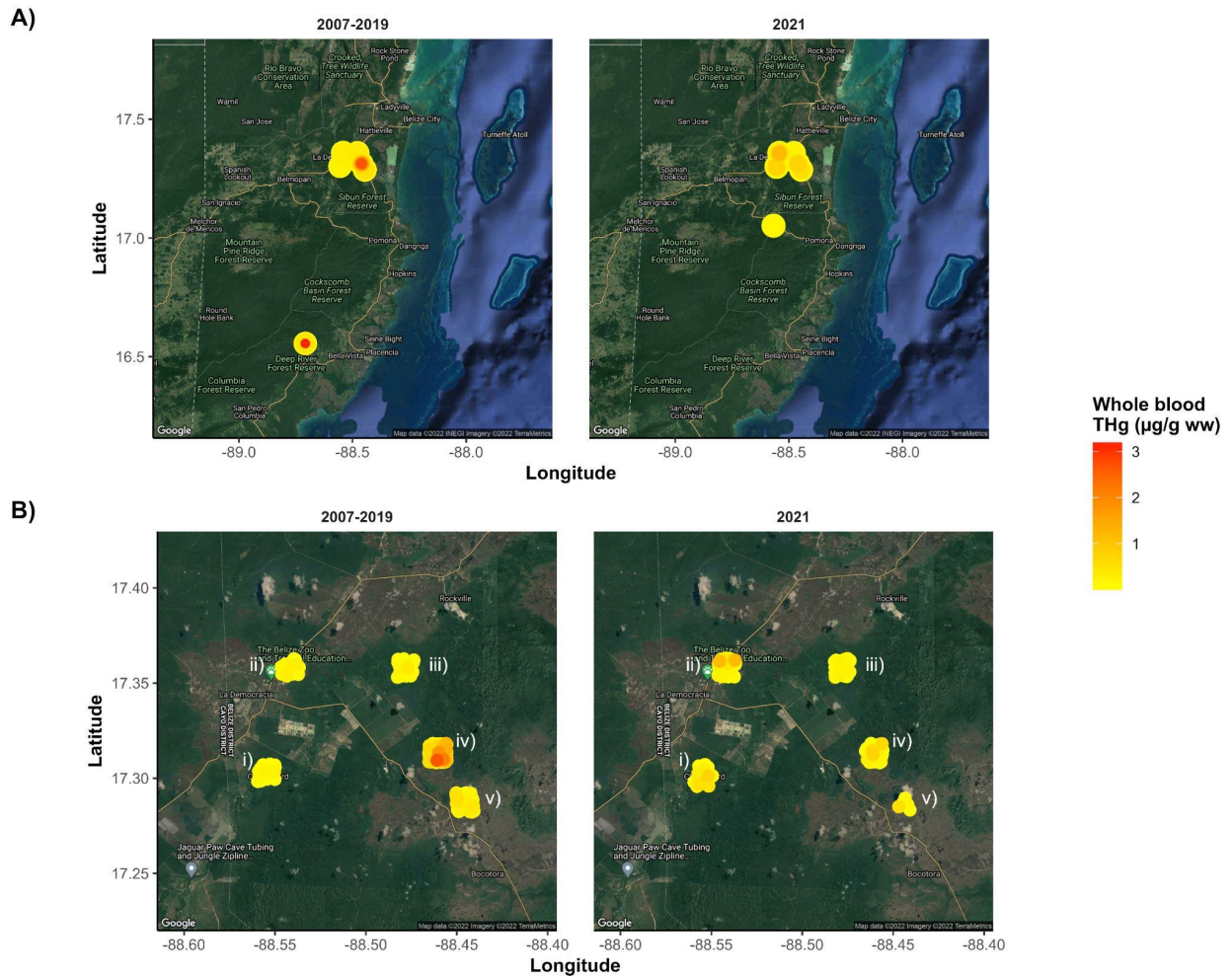


Figure 4. Map displaying total mercury (THg) concentrations ($\mu\text{g/g}$) of 546 whole blood samples from 86 bird species collected throughout central Belize collected from 2007–2021. Map **A)** displays THg concentrations on a country aggregate, while **B)** displays THg concentrations from sites i) Monkey Bay Wildlife Sanctuary, ii) Tropical Education Center, iii) Sibun River, iv) Site B, and v) Heron Roost.

Strengthening scientific collaborations

From September 27–October 1 and October 11–15, 2021, field biologists from University of Belize Environmental Research Institute, BRI, Runaway Creek Nature Reserve, Monkey Bay Wildlife Sanctuary, Belize Audubon Society, and Toledo Institute for Development and Environment (TIDE) convened at TREES for two overnight workshops to learn hands-on bird banding skills while operating an established Monitoreo de Supervivencia Invernal (MoSI, see <https://www.birdpop.org/pages/mosi.php>) station (Fig. 5). This event series provided the perfect venue to unite the majority of the Belizean bird banding community, communicate essential bird

banding theory and field skills, and convey the mission and principal findings of BRI's ongoing Neotropical Hg research. Specifically, as part of his role as workshop co-instructor, Christopher Sayers gave multiple presentations on BRI's previous bird sampling throughout Central America and the Caribbean — highlighting a newly-developed collaborative opportunity for attending organizations.

To overcome previous methodological inconsistencies and inferential uncertainties in the field of tropical ecotoxicology, BRI has developed an international, collaborative data-sharing platform known as the Tropical Research for Avian Conservation and Ecotoxicology (TRACE) Initiative (see <https://www.briwildlife.org/TRACE>). Through the equitable transfer of funding, equipment, data, and authorship among collaborators, TRACE seeks to better inform biodiversity conservation efforts through the understanding of the prevalence, spatiotemporal distribution, toxicokinetics, and biogeochemistry of Hg throughout the tropics — using birds as a convenient, ecologically important, and charismatic proxy. Such standardized information can then be used for long-term Hg monitoring purposes to inform local, national, and international government entities, including Parties of the United Nations Minamata Convention on Mercury.

During the cumulative 10 workshop days, leaders and participants assisted Christopher in collecting over 150 blood and feather samples from resident and migratory bird species. As a result, attendees gained the tissue extraction skills necessary to contribute samples to the TRACE Initiative, as well as pursue research of their own design. As a direct result of this workshop series, BRI is actively collaborating with Foundation for Wildlife Conservation, Monkey Bay Wildlife Sanctuary, University of Belize Environmental Research Institute, Belize Audubon Society, TIDE, Corozal Sustainable Future Initiative (CSFI), and TREES to collect avian tissue samples for later Hg analysis via the TRACE Initiative. Bird Hg samples collected at their respective reserve systems will greatly expand our knowledge about the spatial distribution of Hg throughout Belize, and aid us in identifying biological hotspots across different habitats.



Figure 5. Group pictures of bird Hg sampling training held at Toucan Ridge Ecology and Education Society.

Conclusions

Given Belize's interest in ratifying the United Nations Minamata Convention on Mercury and membership of the newly organized Caribbean Region Mercury Monitoring Network (see <https://briwildlife.org/hgcenter/crmmn/>; Evers and Burton 2021), Belize is uniquely positioned throughout the Neotropics to serve as a long-term study area for assessing the spatiotemporal trends and impacts of Hg exposure to Neotropical birds. Study sites including the Runaway Creek Nature Reserve and TREES provide an ideal framework to build local capacity on bird banding, population science, and ecotoxicological sampling via their educational outreach programs, MoSI stations, and newly-adopted TRACE Initiative.

Recommendations

1. Continue monitoring avian Hg exposure at Runaway Creek Nature Reserve, Tropical Education Center, Monkey Bay Wildlife Sanctuary, and Toucan Ridge Ecology and Education Society as long-term study sites.
2. Expand avian Hg biomonitoring efforts to incorporate sites within the Chiquibul/Maya Mountains to assess potential biotic impacts of artisanal gold mining Hg emissions.
3. Prioritize sampling abundant and widespread invertivore and piscivore sentinel species that exhibit elevated Hg concentrations, including American Pygmy Kingfisher (*Chloroceryle aenea*), Hooded Warbler (*Setophaga citrina*), and Tawny-winged Woodcreeper (*Dendrocincla anabatina*), to broaden the taxonomic and geographic applicability of Hg data from Belize (Appendix 3).
4. Approach Hg sampling using a “food web strategy” by taking soil, leaf, ant, and arachnid samples to better understand avian Hg biomagnification dynamics using stable Hg isotope signatures.

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Figure 6. Barred Antshrike (*Thamnophilus doliatus*) at Runaway Creek Nature Reserve.

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Appendix 1. Total mercury (THg) concentrations ($\mu\text{g/g}$) among all 7 sites sampled throughout Belize from 2007–2021. Number of species sampled, sample size (n), arithmetic mean \pm standard deviation (SD), range, and coefficient of variation (CV) are summarized by site, habitat, and year. A dash (–) indicates there are no data to report.

| Site (latitude, longitude) | Habitat | Year | Species sampled | n | Arithmetic mean \pm SD | Range | CV |
|--|--|-----------|-----------------|-----|--------------------------|-------------|--------|
| Belize Foundation for Research & Environmental Education (16.55578, -88.70777) | Shade coffee/cacao plantation; secondary forest | 2007–2019 | 13 | 65 | 0.155 \pm 0.402 | 0.010–3.195 | 258.9% |
| Monkey Bay Wildlife Sanctuary (17.30258, -88.55478) | River-edge forest; riparian thickets | 2007–2019 | 11 | 28 | 0.066 \pm 0.060 | 0.008–0.222 | 90.4% |
| | | 2021 | 12 | 19 | 0.201 \pm 0.219 | 0.001–0.741 | 108.9% |
| Runaway Creek Nature Reserve--Heron Roost (17.287262, -88.44553) | Riparian thickets; secondary forest | 2007–2019 | 11 | 54 | 0.118 \pm 0.134 | 0.005–0.57 | 112.9% |
| | | 2021 | 5 | 5 | 0.179 \pm 0.202 | 0.004–0.525 | 113.1% |
| Runaway Creek Nature Reserve--Sibun River (17.358047, -88.47801) | Tropical lowland evergreen forest; second-growth shrub | 2007–2019 | 18 | 32 | 0.060 \pm 0.068 | 0.005–0.37 | 114.3% |
| | | 2021 | 17 | 39 | 0.083 \pm 0.070 | 0.003–0.234 | 83.9% |
| Runaway Creek Nature Reserve--Site B (17.313222, -88.460329) | Secondary forest; low, seasonally wet grassland | 2007–2019 | 49 | 150 | 0.259 \pm 0.379 | 0.003–2.659 | 146.4% |
| | | 2021 | 25 | 41 | 0.116 \pm 0.176 | 0.002–0.719 | 151.9% |
| Toucan Ridge Ecology and Education Society (17.05231, -88.56773) | Secondary forest; tropical lowland evergreen forest | 2021 | 14 | 21 | 0.040 \pm 0.041 | 0.002–0.138 | 101.8% |
| Tropical Education Center (17.35812, -88.54162) | Secondary forest; pine-oak forest; gallery forest | 2007–2019 | 19 | 35 | 0.062 \pm 0.081 | 0.001–0.293 | 129.5% |
| | | 2021 | 23 | 57 | 0.150 \pm 0.232 | 0.001–1.187 | 154.9% |

Appendix 2. Whole blood total mercury (THg) concentrations ($\mu\text{g/g}$) among all 86 Neotropical resident and migratory bird species sampled throughout Belize from 2007–2021. Sample size (n), arithmetic mean \pm standard deviation (SD), range, and coefficient of variation (CV) are summarized by family, species, and year. Species are arranged alphabetically by order and family. A dash (–) indicates there are no data to report.

| Family | Common name | Latin name | Year | n | Arithmetic mean \pm SD | Range | CV |
|--|---------------------------|----------------------------------|-----------|-------------------|--------------------------|-------------|-------------|
| Alcedinidae (Kingfishers) | American Pygmy Kingfisher | <i>Chloroceryle aenea</i> | 2007–2019 | 14 | 0.523 \pm 0.418 | 79.9% | 0.069–1.37 |
| | | | 2021 | 7 | 0.683 \pm 0.386 | 56.4% | 0.159–1.187 |
| Alcedinidae (Kingfishers) | Green Kingfisher | <i>Chloroceryle americana</i> | 2007–2019 | 1 | 0.515 | — | — |
| | | | 2021 | 3 | 0.446 \pm 0.048 | 10.7% | 0.395–0.49 |
| Caprimulgidae (Nightjars and Allies) | Common Pauraque | <i>Nyctidromus albicollis</i> | 2021 | 1 | 0.036 | — | — |
| Cardinalidae (Cardinals and Allies) | Blue Bunting | <i>Cyanocopsa parrellina</i> | 2007–2019 | 1 | 0.045 | — | — |
| | | | 2021 | 1 | 0.007 | — | — |
| | Gray-throated Chat | <i>Granatellus sallaei</i> | 2021 | 1 | 0.14 | — | — |
| | Red-throated Ant-Tanager | <i>Habia fuscicauda</i> | 2007–2019 | 9 | 0.065 \pm 0.102 | 157.3% | 0.017–0.337 |
| 2021 | | | 2 | 0.031 \pm 0.035 | 111.9% | 0.007–0.056 | |
| Columbidae (Pigeons and Doves) | Ruddy Ground Dove | <i>Columbina talpacoti</i> | 2021 | 1 | 0.001 | — | — |
| | White-tipped Dove | <i>Leptotila verreauxi</i> | 2007–2019 | 1 | 0.007 | — | — |
| Corvidae (Crows, Jays, and Magpies) | Brown Jay | <i>Psilorhinus morio</i> | 2007–2019 | 1 | 0.027 | — | — |
| Fringillidae (Finches, Euphonias, and Allies) | Yellow-throated Euphonia | <i>Euphonia hirundinacea</i> | 2007–2019 | 2 | 0.006 \pm 0.002 | 37.0% | 0.005–0.008 |
| Furnariidae (Ovenbirds and Woodcreepers) | Ivory-billed Woodcreeper | <i>Xiphorhynchus flavigaster</i> | 2021 | 1 | 0.156 | — | — |
| | Olivaceous Woodcreeper | <i>Sittasomus griseicapillus</i> | 2007–2019 | 2 | 0.112 \pm 0.045 | 40.5% | 0.08–0.144 |

| | | | | | | | |
|--|-------------------------------|-----------------------------------|-----------|----|---------------|--------|-------------|
| | Plain Xenops | <i>Xenops minutus</i> | 2007–2019 | 1 | 0.045 | — | — |
| | Ruddy Woodcreeper | <i>Dendrocincla homochroa</i> | 2007–2019 | 13 | 0.371 ± 0.122 | 32.9% | 0.222–0.598 |
| | | | 2021 | 1 | 0.525 | — | — |
| | Rufous-breasted Spinetail | <i>Synallaxis erythrothorax</i> | 2007–2019 | 1 | 0.045 | — | — |
| | | | 2021 | 3 | 0.066 ± 0.059 | 89.9% | 0.01–0.128 |
| | Streak-headed Woodcreeper | <i>Lepidocolaptes souleyetii</i> | 2007–2019 | 2 | 0.145 ± 0.017 | 11.8% | 0.133–0.157 |
| | Tawny-winged Woodcreeper | <i>Dendrocincla anabatina</i> | 2007–2019 | 11 | 0.261 ± 0.176 | 67.4% | 0.073–0.607 |
| | | | 2021 | 9 | 0.15 ± 0.053 | 35.3% | 0.06–0.234 |
| | Wedge-billed Woodcreeper | <i>Glyphorhynchus spirurus</i> | 2007–2019 | 1 | 0.278 | — | — |
| Galbulidae (Jacamars) | Rufous-tailed Jacamar | <i>Galbula ruficauda</i> | 2021 | 1 | 0.063 | — | — |
| Hirundinidae (Swallows) | Northern Rough-winged Swallow | <i>Stelgidopteryx serripennis</i> | 2007–2019 | 2 | 0.173 ± 0 | 0.1% | 0.173–0.173 |
| Icteridae (Troupials and Allies) | Baltimore Oriole | <i>Icterus galbula</i> | 2007–2019 | 1 | 0.008 | — | — |
| | Black-cowled Oriole | <i>Icterus prothemelas</i> | 2021 | 1 | 0.007 | — | — |
| | Yellow-billed Caciue | <i>Amblycercus holosericeus</i> | 2007–2019 | 3 | 0.112 ± 0.127 | 113.9% | 0.032–0.259 |
| | | | 2021 | 4 | 0.028 ± 0.013 | 46.1% | 0.016–0.045 |
| Icteriidae (Yellow-breasted Chat) | Yellow-breasted Chat | <i>Icteria virens</i> | 2007–2019 | 3 | 0.046 ± 0.065 | 142.3% | 0.006–0.12 |
| | | | 2021 | 1 | 0.059 | — | — |
| Mimidae (Mockingbirds and Thrashers) | Gray Catbird | <i>Dumetella carolinensis</i> | 2007–2019 | 9 | 0.022 ± 0.012 | 52.1% | 0.012–0.048 |
| Momotidae (Motmots) | Lesson's Motmot | <i>Momotus lessonii</i> | 2021 | 1 | 0.036 | — | — |
| Oxyruncidae (Sharpbill, Royal Flycatcher, and Allies) | Royal Flycatcher | <i>Onychorhynchus coronatus</i> | 2007–2019 | 2 | 0.545 ± 0.164 | 30.2% | 0.429–0.662 |

| | | | | | | | |
|---|--------------------------------|--------------------------------|-----------|---------------|---------------|-------------|-------------|
| | | | 2021 | 1 | 0.014 | — | — |
| Parulidae (New World Warblers) | American Redstart | <i>Setophaga ruticilla</i> | 2007–2019 | 4 | 0.263 ± 0.133 | 50.4% | 0.101–0.414 |
| | Black-and-white Warbler | <i>Mniotilta varia</i> | 2007–2019 | 4 | 0.085 ± 0.041 | 48.3% | 0.033–0.13 |
| | | | 2021 | 6 | 0.09 ± 0.018 | 20.2% | 0.07–0.115 |
| | Common Yellowthroat | <i>Geothlypis trichas</i> | 2007–2019 | 14 | 0.127 ± 0.091 | 71.3% | 0.023–0.4 |
| | | | 2021 | 1 | 0.015 | — | — |
| | Hooded Warbler | <i>Setophaga citrina</i> | 2007–2019 | 14 | 0.601 ± 0.819 | 136.2% | 0.047–2.659 |
| | | | 2021 | 12 | 0.153 ± 0.071 | 46.2% | 0.067–0.288 |
| | Kentucky Warbler | <i>Geothlypis formosa</i> | 2007–2019 | 19 | 0.037 ± 0.017 | 46.1% | 0.014–0.085 |
| | | | 2021 | 4 | 0.037 ± 0.012 | 31.9% | 0.023–0.049 |
| | Louisiana Waterthrush | <i>Parkesia motacilla</i> | 2007–2019 | 2 | 0.399 ± 0.001 | 0.4% | 0.398–0.4 |
| | | | 2021 | 6 | 0.297 ± 0.114 | 38.5% | 0.131–0.468 |
| | Magnolia Warbler | <i>Setophaga magnolia</i> | 2007–2019 | 7 | 0.096 ± 0.055 | 57.6% | 0.041–0.208 |
| | Mourning Warbler | <i>Geothlypis philadelphia</i> | 2021 | 1 | 0.028 | — | — |
| | Northern Waterthrush | <i>Parkesia noveboracensis</i> | 2007–2019 | 36 | 0.305 ± 0.33 | 108.2% | 0.047–1.804 |
| | | | 2021 | 6 | 0.166 ± 0.102 | 61.5% | 0.041–0.339 |
| | Ovenbird | <i>Seiurus aurocapilla</i> | 2007–2019 | 7 | 0.093 ± 0.073 | 79.1% | 0.021–0.202 |
| | | | 2021 | 5 | 0.094 ± 0.062 | 65.6% | 0.045–0.201 |
| | Prothonotary Warbler | <i>Protonotaria citrea</i> | 2007–2019 | 6 | 0.609 ± 1.267 | 207.9% | 0.043–3.195 |
| | | 2021 | 5 | 0.185 ± 0.263 | 142.1% | 0.04–0.654 | |
| Swainson's Warbler | <i>Limnothlypis swainsonii</i> | 2021 | 1 | 0.208 | — | — | |
| Tennessee Warbler | <i>Leiothlypis peregrina</i> | 2007–2019 | 1 | 0.022 | — | — | |
| Worm-eating Warbler | <i>Helmitheros vermivorum</i> | 2007–2019 | 2 | 0.022 ± 0.005 | 21.7% | 0.019–0.025 | |

| | | | | | | | |
|---|------------------------|---------------------------------|-----------|----|---------------|--------|-------------|
| | | | 2021 | 3 | 0.054 ± 0.013 | 23.9% | 0.044–0.068 |
| | Yellow Warbler | <i>Setophaga petechia</i> | 2007–2019 | 4 | 0.021 ± 0.008 | 39.4% | 0.013–0.032 |
| Passerellidae (New World Sparrows) | Green-backed Sparrow | <i>Arremonops chloronotus</i> | 2007–2019 | 3 | 0.076 ± 0.041 | 53.9% | 0.037–0.119 |
| | | | 2021 | 14 | 0.028 ± 0.025 | 90.7% | 0.003–0.093 |
| Picidae (Woodpeckers) | Yucatan Woodpecker | <i>Melanerpes pygmaeus</i> | 2021 | 2 | 0.009 ± 0.004 | 47.5% | 0.006–0.013 |
| Pipridae (Manakins) | Red-capped Manakin | <i>Ceratopipra mentalis</i> | 2007–2019 | 12 | 0.013 ± 0.005 | 38.3% | 0.006–0.024 |
| | | | 2021 | 4 | 0.003 ± 0.001 | 27.5% | 0.003–0.005 |
| | White-collared Manakin | <i>Manacus candei</i> | 2007–2019 | 5 | 0.024 ± 0.032 | 134.1% | 0.005–0.08 |
| | | | 2021 | 5 | 0.003 ± 0.001 | 45.7% | 0.001–0.004 |
| Poliptilidae (Gnatcatchers) | Long-billed Gnatwren | <i>Ramphocaenus melanurus</i> | 2007–2019 | 3 | 0.077 ± 0.055 | 71.8% | 0.04–0.141 |
| | | | 2021 | 1 | 0.016 | — | — |
| Scolopacidae (Sandpipers and Allies) | Spotted Sandpiper | <i>Actitis macularius</i> | 2021 | 1 | 0.741 | — | — |
| Thamnophilidae (Typical Antbirds) | Barred Antshrike | <i>Thamnophilus doliatus</i> | 2007–2019 | 1 | 0.167 | — | — |
| | | | 2021 | 2 | 0.032 ± 0.031 | 96.5% | 0.01–0.054 |
| | Dusky Antbird | <i>Cercomacroides tyrannina</i> | 2007–2019 | 8 | 0.046 ± 0.014 | 29.1% | 0.026–0.059 |
| | | | 2021 | 2 | 0.054 ± 0.058 | 107.7% | 0.013–0.095 |
| | Great Antshrike | <i>Taraba major</i> | 2021 | 1 | 0.09 | — | — |
| Thraupidae (Tanagers and Allies) | Buff-throated Saltator | <i>Saltator maximus</i> | 2007–2019 | 1 | 0.001 | — | — |
| | Gray-headed Tanager | <i>Eucometis penicillata</i> | 2007–2019 | 15 | 0.251 ± 0.119 | 47.2% | 0.101–0.509 |
| | | | 2021 | 9 | 0.154 ± 0.059 | 38.5% | 0.083–0.281 |
| | Morelet's Seedeater | <i>Sporophila moreletii</i> | 2007–2019 | 15 | 0.017 ± 0.021 | 123.7% | 0.006–0.091 |

| | | | | | | | |
|--|--------------------------|----------------------------------|-----------|----|---------------|--------|-------------|
| | Thick-billed Seed-Finch | <i>Sporophila funerea</i> | 2007–2019 | 1 | 0.007 | — | — |
| Tityridae (Tityras and Allies) | Northern Schiffornis | <i>Schiffornis veraepacis</i> | 2007–2019 | 1 | 0.003 | — | — |
| | | | 2021 | 1 | 0.002 | — | — |
| | Rose-throated Becard | <i>Pachyramphus aglaiae</i> | 2007–2019 | 1 | 0.05 | — | — |
| Troglodytidae (Wrens) | Spot-breasted Wren | <i>Pheugopedius maculipectus</i> | 2007–2019 | 12 | 0.108 ± 0.055 | 51.5% | 0.027–0.187 |
| | | | 2021 | 11 | 0.053 ± 0.081 | 152.9% | 0.009–0.289 |
| | White-bellied Wren | <i>Uropsila leucogastra</i> | 2007–2019 | 2 | 0.224 ± 0.023 | 10.3% | 0.208–0.24 |
| | | | 2021 | 2 | 0.074 ± 0.066 | 89.0% | 0.028–0.121 |
| | White-breasted Wood-Wren | <i>Henicorhina leucosticta</i> | 2007–2019 | 7 | 0.194 ± 0.316 | 163.2% | 0.053–0.909 |
| | | | 2021 | 1 | 0.28 | — | — |
| Turdidae (Thrushes and Allies) | Clay-colored Thrush | <i>Turdus grayi</i> | 2021 | 13 | 0.021 ± 0.022 | 104.3% | 0.005–0.079 |
| | | | | | | | |
| | Swainson's Thrush | <i>Catharus ustulatus</i> | 2021 | 5 | 0.004 ± 0.002 | 44.7% | 0.002–0.007 |
| | Veery | <i>Catharus fuscescens</i> | 2021 | 1 | 0.025 | — | — |
| | White-throated Thrush | <i>Turdus assimilis</i> | 2007–2019 | 1 | 0.038 | — | — |
| | Wood Thrush | <i>Hylocichla mustelina</i> | 2007–2019 | 3 | 0.084 ± 0.032 | 37.8% | 0.054–0.117 |
| Tyrannidae (Tyrant Flycatchers) | Acadian Flycatcher | <i>Empidonax virescens</i> | 2007–2019 | 1 | 0.147 | — | — |
| | | | 2021 | 4 | 0.178 ± 0.036 | 20.0% | 0.136–0.216 |
| | Bright-rumped Attila | <i>Attila spadiceus</i> | 2007–2019 | 1 | 0.107 | — | — |
| | | | 2021 | 2 | 0.051 ± 0.037 | 71.5% | 0.025–0.077 |
| | Dusky-capped Flycatcher | <i>Myiarchus tuberculifer</i> | 2007–2019 | 1 | 0.031 | — | — |
| | | | 2021 | 2 | 0.039 ± 0 | 1.2% | 0.038–0.039 |
| | Eastern Wood-Pewee | <i>Contopus virens</i> | 2021 | 1 | 0.484 | — | — |

| | | | | | | | |
|---|----------------------------|-----------------------------------|-----------|----|---------------|--------|-------------|
| | Eye-ringed Flatbill | <i>Rhynchocyclus brevirostris</i> | 2007–2019 | 2 | 0.066 ± 0.042 | 64.3% | 0.036–0.096 |
| | Great Kiskadee | <i>Pitangus sulphuratus</i> | 2021 | 2 | 0.088 ± 0.001 | 0.9% | 0.087–0.088 |
| | Greenish Elaenia | <i>Myiopagis viridicata</i> | 2007–2019 | 1 | 0.02 | — | — |
| | Northern Bentbill | <i>Oncostoma cinereigulare</i> | 2007–2019 | 2 | 0.107 ± 0.009 | 8.3% | 0.1–0.113 |
| | | | 2021 | 3 | 0.044 ± 0.002 | 4.5% | 0.041–0.045 |
| | Ochre-bellied Flycatcher | <i>Mionectes oleagineus</i> | 2007–2019 | 28 | 0.031 ± 0.05 | 162.8% | 0.009–0.278 |
| | Social Flycatcher | <i>Myiozetetes similis</i> | 2007–2019 | 1 | 0.017 | — | — |
| | Sulphur-bellied Flycatcher | <i>Myiodynastes luteiventris</i> | 2007–2019 | 2 | 0.057 ± 0.023 | 39.7% | 0.041–0.073 |
| | Tropical Kingbird | <i>Tyrannus melancholicus</i> | 2007–2019 | 1 | 0.042 | — | — |
| | Tropical Pewee | <i>Contopus cinereus</i> | 2007–2019 | 3 | 0.174 ± 0.042 | 23.96% | 0.126–0.203 |
| | Yellow-bellied Elaenia | <i>Elaenia flavogaster</i> | 2007–2019 | 3 | 0.007 ± 0.003 | 33.9% | 0.005–0.01 |
| | Yellow-bellied Flycatcher | <i>Empidonax flaviventris</i> | 2007–2019 | 1 | 0.101 | — | — |
| | | | 2021 | 1 | 0.126 | — | — |
| | Yellow-olive Flycatcher | <i>Tolmomyias sulphurescens</i> | 2007–2019 | 1 | 0.066 | — | — |
| Vireonidae (Vireos, Shrike-Babblers, and Erpornis) | Lesser Greenlet | <i>Pachysylvia decurtata</i> | 2021 | 1 | 0.084 | — | — |
| | Mangrove Vireo | <i>Vireo pallens</i> | 2007–2019 | 7 | 0.173 ± 0.107 | 61.7% | 0.073–0.315 |
| | | | 2021 | 1 | 0.019 | — | — |
| | Red-eyed Vireo | <i>Vireo olivaceus</i> | 2007–2019 | 2 | 0.025 ± 0.0 | 0.0% | 0.025–0.025 |
| | White-eyed Vireo | <i>Vireo griseus</i> | 2007–2019 | 3 | 0.057 ± 0.038 | 66.8% | 0.018–0.093 |
| | Yellow-green Vireo | <i>Vireo flavoviridis</i> | 2007–2019 | 13 | 0.02 ± 0.008 | 40.3% | 0.009–0.035 |
| | | | 2021 | 1 | 0.007 | — | — |

Appendix 3. Summary of 75 abundant, widespread species with relatively high trophic positions that could be useful sentinels for assessing avian MeHg exposure in various Neotropical habitats. An asterisks (*) indicates that other members within this group may also be useful. Relative abundance was evaluated by visually inspecting mist-netting capture data and country-level eBird Bar Charts (<https://ebird.org/GuideMe?cmd=changeLocation>). The exclusion of a family or species from this list does not necessarily signify that those taxa would be poor MeHg bioindicators, but that sample acquisition could be challenging. Focal species selection for a given research effort should ultimately be informed by their abundance at the intended sampling locations, their diet and trophic niche, as well as the geographic scale of interest. Species are arranged alphabetically by order and family. Species highlighted in green may be especially useful due to their widespread range.

| Order | Family | Common name | Latin name | Trophic niche | Migratory status | Primary habitat(s) | Region(s) occupied |
|-------------------------|---|---------------------------|-------------------------------|------------------|------------------|--|----------------------------------|
| Caprimulgiformes | Caprimulgidae (Nightjars and Allies) * | Common Pauraque | <i>Nyctidromus albicollis</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest; tropical deciduous forest | Texas, USA to Argentina |
| | Trochilidae (Hummingbirds) * | Rufous-tailed Hummingbird | <i>Amazilia tzacatl</i> | Nectarivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Ecuador |
| | | White-necked Jacobin | <i>Florisuga mellivora</i> | Nectarivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Bolivia |
| | (Sandpipers and Allies) * | Spotted Sandpiper | <i>Actitis macularia</i> | Aquatic predator | Full migrant | Riverine sand beaches; freshwater lakes and ponds | Canada to Argentina; West Indies |
| Coraciiformes * | Alcedinidae (Kingfishers) * | American Pygmy Kingfisher | <i>Chloroceryle aenea</i> | Aquatic predator | Resident | Streams; freshwater lakes and ponds | Mexico to Paraguay |
| | | Green Kingfisher | <i>Chloroceryle americana</i> | Aquatic predator | Resident | Streams; freshwater lakes and ponds | Texas, USA to Argentina |
| | Momotidae (Motmots) * | Lesson's Motmot | <i>Momotus lessonii</i> | Omnivore | Resident | Tropical lowland evergreen forest; montane evergreen forest; secondary | Mexico to Panama |

| | | | | | | | |
|-----------------------|---|--------------------------|----------------------------------|-------------|-----------------|--|-----------------------------------|
| | | | | | | forest | |
| Cuculiformes * | Cuculidae (Cuckoos) * | Groove-billed Ani | <i>Crotophaga sulcirostris</i> | Invertivore | Resident | Second-growth scrub; riparian thickets | Texas, USA to Peru |
| | | Squirrel Cuckoo | <i>Piaya cayana</i> | Invertivore | Resident | Tropical lowland evergreen forest; tropical deciduous forest; secondary forest | Mexico to Uruguay |
| Galbuliformes | Galbulidae (Jacamars) * | Rufous-tailed Jacamar | <i>Galbula ruficauda</i> | Invertivore | Resident | Tropical lowland evergreen forest; gallery forest | Mexico to Brazil |
| Passeriformes | Cardinalidae (Cardinals and Allies) * | Hepatic Tanager | <i>Piranga flava</i> | Invertivore | Resident | Pine-oak forest; gallery forest; tropical deciduous forest | Colorado, USA to Argentina |
| | | Red-crowned Ant-Tanager | <i>Habia rubica</i> | Invertivore | Resident | Tropical lowland evergreen forest | Mexico to Brazil |
| | | Red-throated Ant-Tanager | <i>Habia fuscicauda</i> | Invertivore | Resident | Tropical lowland evergreen forest; flooded tropical evergreen forest | Mexico to Colombia |
| | | Summer Tanager | <i>Piranga rubra</i> | Invertivore | Partial migrant | Gallery forest | Iowa, USA to Bolivia; West Indies |
| | Formicariidae (Antthrushes) * | Mayan Anthrush | <i>Formicarius moniliger</i> | Invertivore | Resident | Tropical lowland evergreen forest; flooded tropical evergreen forest | Mexico to Honduras |
| | Furnariidae (Ovenbirds and Woodcreepers) * | Ivory-billed Woodcreeper | <i>Xiphorhynchus flavigaster</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest; tropical deciduous forest | Mexico to Costa Rica |
| | | Olivaceous Woodcreeper | <i>Sittasomus griseicapillus</i> | Invertivore | Resident | Tropical lowland evergreen | Mexico to Uruguay |

| | | | | | | |
|--|-------------------------------|-----------------------------------|-------------|-----------------|--|-------------------------------|
| | Plain Xenops | <i>Xenops minutus</i> | Invertivore | Resident | forest; montane evergreen forest; tropical deciduous forest | |
| | | | | | Tropical lowland evergreen forest; flooded tropical evergreen forest | Mexico to Brazil |
| | Ruddy Woodcreeper | <i>Dendrocincla homochroa</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest; tropical deciduous forest | Mexico to Venezuela |
| | Tawny-winged Woodcreeper | <i>Dendrocincla anabatina</i> | Invertivore | Resident | Tropical lowland evergreen forest | Mexico to Panama |
| | Wedge-billed Woodcreeper | <i>Glyphorhynchus spirurus</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest | Mexico to Brazil |
| Hirundinidae (Swallows) * | Barn Swallow | <i>Hirundo rustica</i> | Invertivore | Partial migrant | Pastures/agricultural lands; Northern temperate grassland | Worldwide |
| | Gray-breasted Martin | <i>Progne chalybea</i> | Invertivore | Partial migrant | Second-growth scrub; pastures/agricultural lands | Mexico to Argentina |
| | Northern Rough-winged Swallow | <i>Stelgidopteryx serripennis</i> | Invertivore | Partial migrant | Second-growth scrub; pastures/agricultural lands | Canada to Panama; West Indies |
| Oxyruncidae (Sharpbill, Royal Flycatcher, and Allies) * | Royal Flycatcher | <i>Onychorhynchus coronatus</i> | Invertivore | Resident | Tropical lowland evergreen forest | Mexico to Brazil |
| Parulidae (New World Warblers) * | American Redstart | <i>Setophaga ruticilla</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; montane evergreen forest | Canada to Peru; West Indies |

| | | | | | | |
|---|---------------------------|---------------------------------|-------------|-----------------|--|-------------------------------------|
| | Black-and-white Warbler | <i>Mniotilta varia</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; montane evergreen forest | Canada to Ecuador; West Indies |
| | Common Yellowthroat | <i>Geothlypis trichas</i> | Invertivore | Partial migrant | Freshwater/saltwater/brackish marshes | Canada to Panama; West Indies |
| | Golden-crowned Warbler | <i>Basileuterus culicivorus</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest | Mexico to Argentina |
| | Gray-crowned Yellowthroat | <i>Geothlypis poliocephala</i> | Invertivore | Resident | Second-growth scrub; riparian thickets | Mexico to Panama |
| | Hooded Warbler | <i>Setophaga citrina</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; secondary forest | Canada to Venezuela; West Indies |
| | Northern Waterthrush | <i>Parkesia noveboracensis</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; secondary forest | Alaska, USA to Ecuador; West Indies |
| | Yellow Warbler | <i>Setophaga petechia</i> | Invertivore | Partial migrant | Gallery forest; secondary forest; mangrove forest | Alaska, USA to Peru; West Indies |
| Passerellidae (New World Sparrows) * | Green-backed Sparrow | <i>Arremonops chloronotus</i> | Omnivore | Resident | Tropical lowland evergreen forest; tropical deciduous forest | Mexico to Honduras |
| | Orange-billed Sparrow | <i>Arremon aurantirostris</i> | Omnivore | Resident | Tropical lowland evergreen forest | Mexico to Peru |
| Poliophtilidae (Gnatcatchers) * | Blue-gray Gnatcatcher | <i>Poliophtila caerulea</i> | Invertivore | Partial migrant | Tropical deciduous forest; tropical lowland evergreen forest | Canada to Honduras; West Indies |
| | Tropical Gnatcatcher | <i>Poliophtila plumbea</i> | Invertivore | Resident | Tropical deciduous | Mexico to Brazil |

| | | | | | | |
|--|--------------------------|----------------------------------|-------------|----------|--|----------------------|
| | | | | | forest; tropical lowland evergreen forest; Arid lowland/montane scrub | |
| Thamnophilidae (Typical Antbirds) * | Barred Antshrike | <i>Thamnophilus doliatus</i> | Invertivore | Resident | Second-growth scrub; riparian thickets; river island scrub | Mexico to Argentina |
| | Dot-winged Antwren | <i>Microrhophias quixensis</i> | Invertivore | Resident | Tropical lowland evergreen forest | Mexico to Brazil |
| | Dusky Antbird | <i>Cercomacroides tyrannina</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Brazil |
| | Great Antshrike | <i>Taraba major</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest; riparian thickets | Mexico to Argentina |
| Thraupidae (Tanagers and Allies) * | Gray-headed Tanager | <i>Eucometis penicillata</i> | Invertivore | Resident | Tropical lowland evergreen forest; gallery forest; secondary forest | Mexico to Paraguay |
| Troglodytidae (Wrens) * | Gray-breasted Wood-Wren | <i>Henicorhina leucophrys</i> | Invertivore | Resident | Montane evergreen forest | Mexico to Bolivia |
| | House Wren | <i>Troglodytes aedon</i> | Invertivore | Resident | Second-growth scrub; arid lowland/montane scrub | Canada to Chile |
| | Spot-breasted Wren | <i>Pheugopedius maculipectus</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Costa Rica |
| | White-breasted Wood-Wren | <i>Henicorhina leucosticta</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest | Mexico to Peru |

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| Turdidae (Thrushes and Allies) * | Clay-colored Thrush | <i>Turdus grayi</i> | Omnivore | Resident | Tropical lowland evergreen forest; tropical deciduous forest; secondary forest | Texas, USA to Colombia |
| | Swainson's Thrush | <i>Catharus ustulatus</i> | Invertivore | Full migrant | Montane evergreen forest; tropical lowland evergreen forest | Alaska, USA to Argentina |
| | Wood Thrush | <i>Hylocichla mustelina</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; secondary forest | Canada to Colombia |
| Tyrannidae (Tyrant Flycatchers) * | Boat-billed Flycatcher | <i>Megarynchus pitangua</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Argentina |
| | Bright-rumped Attila | <i>Attila spadiceus</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest; montane evergreen forest | Mexico to Bolivia |
| | Common Tody-Flycatcher | <i>Todirostrum cinereum</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest; tropical deciduous forest | Mexico to Brazil |
| | Dusky-capped Flycatcher | <i>Myiarchus tuberculifer</i> | Invertivore | Resident | Montane evergreen forest; tropical lowland evergreen forest; tropical deciduous forest | Arizona, USA to Argentina |
| | Eastern Wood-Pewee | <i>Contopus virens</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; secondary forest | Canada to Peru; West Indies |

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| | Great Kiskadee | <i>Pitangus sulphuratus</i> | Omnivore | Resident | Secondary forest; riparian thickets | Texas, USA to Argentina |
| | Sepia-capped Flycatcher | <i>Leptopogon amaurocephalus</i> | Invertivore | Resident | Tropical lowland evergreen forest; secondary forest | Mexico to Brazil |
| | Social Flycatcher | <i>Myiozetetes similis</i> | Invertivore | Resident | Tropical lowland evergreen forest; tropical deciduous forest | Mexico to Argentina |
| | Tropical Kingbird | <i>Tyrannus melancholicus</i> | Invertivore | Resident | Secondary forest; gallery forest; river-edge forest | Arizona, USA to Argentina |
| | Tropical Pewee | <i>Contopus cinereus</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest | Mexico to Argentina |
| | Yellow-bellied Elaenia | <i>Elaenia flavogaster</i> | Invertivore | Resident | Second-growth scrub; riparian thickets | Mexico to Brazil |
| | Yellow-olive Flycatcher | <i>Tolmomyias sulphurescens</i> | Invertivore | Resident | Tropical lowland evergreen forest; montane evergreen forest; tropical deciduous forest | Mexico to Uruguay |
| Vireonidae (Vireos, Shrike-Babblers, and Erpornis) * | Lesser Greenlet | <i>Pachysylvia decurtata</i> | Invertivore | Resident | Tropical lowland evergreen forest; tropical deciduous forest | Mexico to Ecuador |
| | Mangrove Vireo | <i>Vireo pallens</i> | Invertivore | Resident | Tropical deciduous forest; mangrove forest | Mexico to Costa Rica |
| | Red-eyed Vireo | <i>Vireo olivaceus</i> | Invertivore | Full migrant | Tropical lowland evergreen forest; secondary forest | Canada to Bolivia |

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| | | Rufous-browed Peppershrike | <i>Cyclarhis gujanensis</i> | Invertivore | Resident | Tropical lowland evergreen forest; gallery forest | Mexico to Argentina |
| | | Tawny-crowned Greenlet | <i>Tunchiornis ochraceiceps</i> | Invertivore | Resident | Tropical lowland evergreen forest | Mexico to Bolivia |
| Pelecaniformes | Ardeidae (Herons, Egrets, and Bitterns) * | Black-crowned Night-Heron | <i>Nycticorax nycticorax</i> | Aquatic predator | Partial migrant | Freshwater/saltwater/brackish marshes; freshwater lakes and ponds; rivers | Worldwide |
| | | Great Egret | <i>Ardea alba</i> | Aquatic predator | Partial migrant | Freshwater/saltwater/brackish marshes; coastal sand beaches/mudflats | Worldwide |
| | | Green Heron | <i>Butorides virescens</i> | Aquatic predator | Partial migrant | Freshwater lakes, ponds, and marshes | United States to Ecuador; West Indies |
| | Pelecanidae (Pelicans) * | Brown Pelican | <i>Pelecanus occidentalis</i> | Aquatic predator | Partial migrant | Coastal waters | United States to Peru; West Indies |
| Podicipediformes * | Podicipedidae (Grebes) * | Least Grebe | <i>Tachybaptus dominicus</i> | Aquatic predator | Resident | Freshwater marshes, lakes, and ponds | Texas, USA to Uruguay |
| | | Pied-billed Grebe | <i>Podilymbus podiceps</i> | Aquatic predator | Resident | Freshwater lakes, ponds, and marshes | Canada to Argentina; West Indies |
| Suliformes | Anhingidae (Anhingas) | Anhinga | <i>Anhinga anhinga</i> | Aquatic predator | Resident | Freshwater lakes and ponds; rivers | United States to Uruguay |
| | Fregatidae (Frigatebirds) | Magnificent Frigatebird | <i>Fregata magnificens</i> | Aquatic predator | Resident | Coastal waters; pelagic waters; mangrove forest | United States to Uruguay; Cape Verde |
| | Phalacrocoracidae (Cormorants and Shags) * | Neotropic Cormorant | <i>Phalacrocorax brasilianus</i> | Aquatic predator | Resident | Coastal waters; freshwater lakes and ponds; rivers | Texas, USA to Chile; Cuba |
| | Sulidae (Boobies and Gannets) * | Brown Booby | <i>Sula leucogaster</i> | Aquatic predator | Resident | Coastal waters; pelagic waters | Worldwide |

